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A New Approach for Utilization of Modified *Epicoccum Nigrum* Biomass by Ferric Chloride Nanoparticles for Biosorption of Thorium

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Abstract

Biosorption of thorium by *Epicoccum nigrum* biomass was determined at varying experimental conditions using a batch technique. High biosorption efficiency was observed for dead biomass 500 mg/50 mL of *E. nigrum* within 60 min of contact, at pH 6.8, temperature of 30 °C and agitation speed of 150 rpm with 150 mg/L as initial thorium concentration. Scanning electron microscope (SEM) equipped with electron dispersive X-ray (EDX) was used for characterization of the biosorbent agent after and before biosorption process to identify composition of elements of the dead biomass after and before biosorption of thorium. Infrared spectroscopy (FTIR) study revealed that the amide groups interact with thorium particles, which was accountable for the stability of thorium complex. This analysis confirmed the presence of proteins as stabilizing and capping agents. Furthermore, X-ray diffraction (XRD) analysis reveals the presence of metallic thorium ions.

Keywords

Epicoccum nigrum, Ferric chloride nanoparticles, Biosorption and thorium

Introduction

There are many methods used for recovering metal ions from pregnant solution such as chemical precipitation, ion exchange, solvent extraction and sorption [1]. These traditional techniques suffer from several disadvantages, such as high consumption of reagent and energy, high operational cost, low selectivity and generation of secondary pollutants [2]. Therefore, alternative techniques for the bioremediation of metal using bacteria, fungi and yeast are currently being investigated. Biosorption can be defined as the removal of metal or metalloid species, particulates and compounds from solution by non-living and living biological materials [3]. The fungus-mediated green technology approach towards the biosorption of metal has several advantages, such as easy and simple scaling up, easy processing, economic viability, biomass handling, and recovery of large surface areas with optimum growth of mycelia [4]. Using dead biomass is more advantageous than the using live biomass as there are no requirements of growth media or nutrients, no toxicity concerns. Also, easy techniques were developed to desorb contaminants from the dead biomass and reuse them [5].

Fungi cell walls contain large quantity of polysaccharides and proteins, which offer many functional groups (such as carboxyl, sulphate, hydroxyl, phosphate and amino groups) for binding metal ions [6]. The reaction between microorganisms and metal ions have been well investigated [7] and the ability of microorganisms to extract and/or biosorb metals is already

used in biotechnological processes such as bioremediation through the biosorption of toxic metals such as thorium. Studies in the literature have reported the use of dead fungal biomass for biosorption of metal ions as nickel by *Hypocrea lixii* [8]. The desorption process has the purpose of recovering the biosorbed metals while regenerating the biomass that could be used in subsequent sorption and desorption cycles [9]. Fungi are frequently used in bioremediation processes since they are able to grow and adapt under extreme conditions of pH, nutrient and temperature availability, as well as at high metal concentrations [10].

Thorium is considered one of the strategic that imparting high strength and creep resistance at elevated temperatures [11]. Thorium Th is a naturally occurring radionuclide with nuclear importance due to its application as an alternative fuel by converting it into 233 U. It is majorly resulted as a strong gamma-emitting by-product during nuclear reactor

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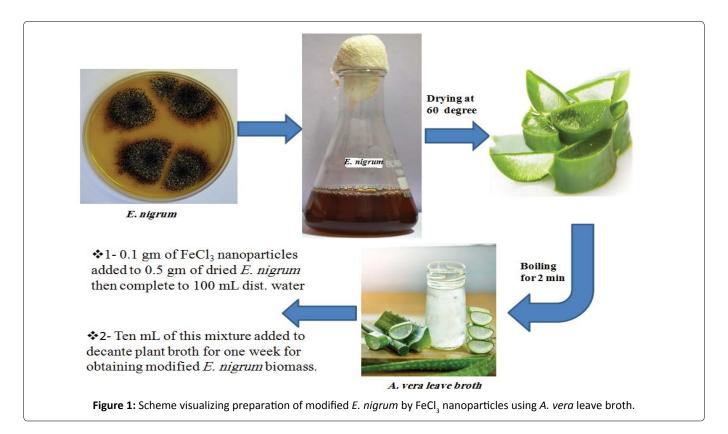
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operations and is found in various industrial effluents [12]. Nanotechnology research deals with manipulation and production of nanoparticles that have new properties, which differ from their corresponding bulk material form. These differences are probably due to effects such as quantum size effect, surface effect, and macro-scopic quantum tunneling [13]. Nanoparticles have been widely studied over the past decade due their application in different areas such as recording media, catalysts, sensors, etc. [14].

This paper reports a novel approach for the modification of the surface area of the fungal by ferric chloride nanoparticles to be used in the uptake of thorium by dead biomass of the filamentous fungus *E. nigrum*. A study of biosorption coupled with process of optimization of the surface area using ferric chloride nanoparticles was conducted. Furthermore, optimization of different physico-chemical parameters, i.e., pH, temperature, initial metal concentration, interaction time and others, were done to ensure efficient uptake of thorium.

Materials and Methods

Microorganism and biosorbent preparation

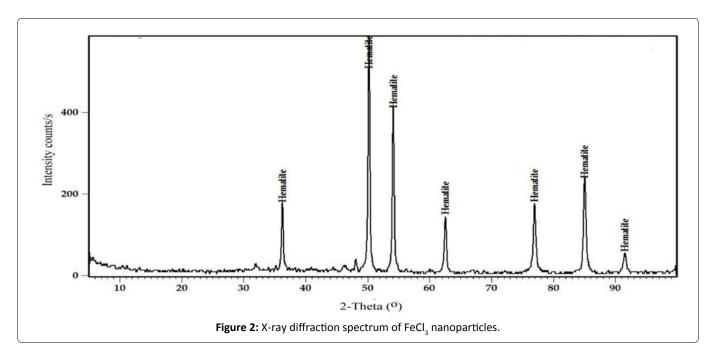
E. nigrum was isolated in Fungal Laboratory of Zagazig University. Modified Czapek's-Dox agar (MCDA) was used for growth of the tested fungi. The medium composition is as follows (g/L): sucrose 30; NaNO $_3$ 3; KH $_2$ PO $_4$ 1; MgSO $_4$.7H $_2$ O 0.5; KCl 0.5; FeSO $_4$ 0.01; yeast extract 10 and agar-agar [15]. The culture was incubated for 15 days at 30°C, and 150 rpm. After 7-8 days of growth, the harvested cells were washed twice with deionized distilled water till the pH of the wash solution was in the near neutral range. Then, it was dried at 60 °C for 24 h before use. The dried mat was ground to obtain uniform sized particles [16].

Preparation of *E. nigrum* surface area using FeCl₃ nanoparticles

Aloe vera leaf broth is used for the pretreatment of the *E. nigrum* surface area by FeCl₃ nanoparticles. Aloe vera leaf broth prepared by taking 20 g of thoroughly washed and finely cut *A. vera* leaves in a 500-mL Erlenmeyer flask with 100 mL of sterile distilled water and then boiling the mixture for 2 min then finally decanting it. For preparation of microorganism 0.1 gm of FeCl₃ nanoparticles added to 0.5 gm of dried *E. nigrum* for homogeneity well then completes the volume to 100 mL with deionized distilled water. Ten mL of this mixture added to decante plant broth for one week. After that, the mixture was filtrated to obtain modified *E. nigrum* surface area attached by FeCl₃ nanoparticles through Aloe vera leaf broth. Then, it was grinded to 125 mm size for biosorption experiments (Figure 1).

Thorium solution preparation and ferric chloride nanoparticles characterization

All Thorium stock solution (1 g/L) was prepared separately from $Th(NO_3)_4$. 4H_2O which obtained from Alfa Aesar. The stock solution was prepared by diluting the stock solution to appropriate volume. Particle size of the used iron chloride nanoparticles is measured using dynamic light scattering technique in a 180° heterodyne-backscatter arrangement using Nanotrac Wave II device. The used nanoparticles show D50 of 100 \pm 10 nanometer in size with PDI of 0.1528 \pm 0.1. Iron chloride nanoparticles were obtained from Zagazig University, Egypt. The obtained X-ray diffraction spectrum appeared distinct peaks indicating the deposition of crystallized hematite. Diffractogram of the hematite displayed maximum number of peaks to iron oxide between



20 of 37.11°, 20 of 50.37°, 20 of 54.56°, 20 of 64.34°, 20 of 77.35°, 20 of 86.49° and 20 of 29.56° as illustrated in Figure 2.

Factors affecting the biosorption process

Biosorption experiments were conducted with 50 mL of 50 mg/L initial concentration of Th(IV) ions solution in 250 mL Erlenmeyer flask for 60 minutes contact time at room temperature 26 ± 1 °C. The effect of initial thorium concentration (50-500 mg/L), pH (2-6), biosorbent dose (0.1-1.5 g), contact time (10-120 min), temperature (20-60 °C) and agitation speed (50-250 rpm) on the biosorption of thorium was analyzed. These experiments were carried out at the desired pH, temperature, initial metal concentration, contact time, agitation speed and biosorbent dose using 50 mL of a 100 mg/L Th(IV) ions solution.

The supernatants were filtered through a (Whatman No. 41) filter paper and the concentration of Th(IV)in the solution was measured before and after equilibrium by (MetertechIno model Sp-5001, UV-Visible spectrophotometer) using arsenazoIII [17]. The efficiency (R) of metal removal was calculated using the following equation:

$$R = \frac{Co - Ce}{Co} x 100$$

Where Co and Ce are initial and equilibrium metal concentrations, respectively.

Characterization of E. nigrum biomass surface area after pretreatment and biosorption of thorium

The change in the surface morphology of pore structure, treated and loaded biomass of *E. nigrum* were observed under scanning electron microscope (SEM), Philips XL30 vacuum at 30 KV.SEM equipped with energy dispersive X-ray (EDX) to identify the composition of elements of the tested biomasses of *E. nigrum*. The apparatus was presented in nuclear materials authority (NMA), Cairo, Egypt.

Infrared spectrophotometer was used to determine the functional groups presented on the surface of *E. nigrum* biomass before and after loading with Th(IV). The samples were prepared as KBr discs. The apparatus used was (Model JASCO 4100 FTIR spectrophotometer, Japan) which presented in the Micro Analytical Center, Cairo University, Egypt.

X-ray Diffraction (XRD) patterns of tested samples were recorded using (PHILIPS PW 3710/31 diffractometer with automatic sample changer PW 1775, (21 positions), Scintillation counter, Cu-target tube and Ni filter at 40 Kv and 30 Am. This instrument is connecting to a computer system using X-40 diffraction program and ASTM cards for mineral identification. The apparatus was found in (NMA), Cairo, Egypt.

Desorption study

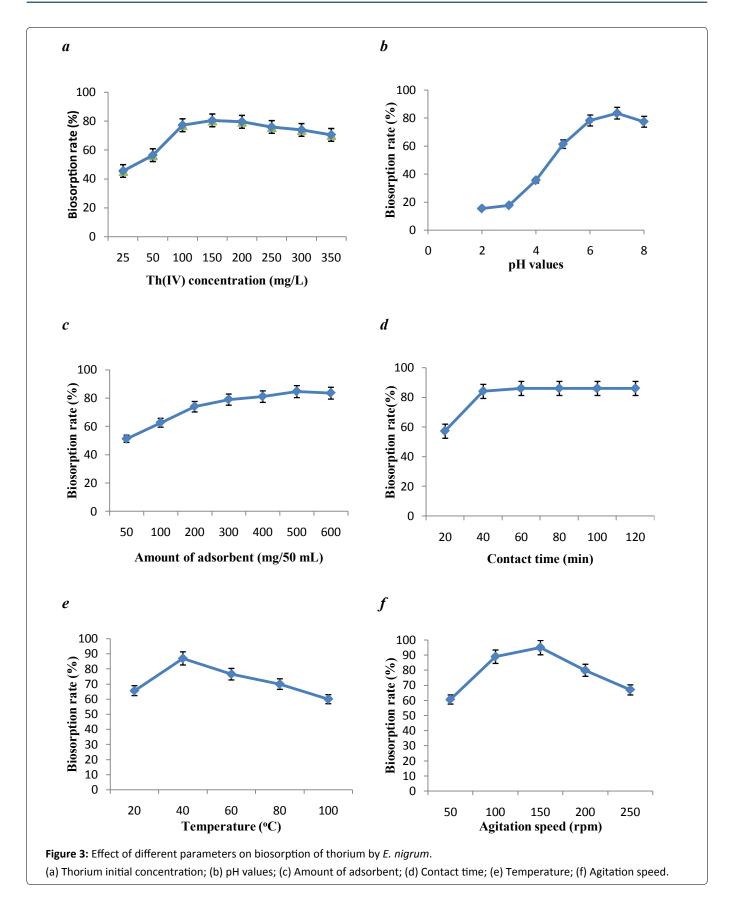
Desorption experiments were performed by mixing *E. nigrum* loaded biomass (500 mg dry weight) with desorbing agents 50 mL of (0.1, 0.2, 0.3 N HCl, 0.1, 0.2, 0.3 N HNO $_3$ and 0.1, 0.2, 0.3 N H $_2$ SO $_4$) for 1 h on a rotary shaker (175 rpm) at 30 °C for three cycles to study the elution efficiency [18].

Results and Discussion

The fungus *E. nigrum* was selected to investigate the uptake of thorium from aqueous medium. Some fungi are recognized as hyperaccumulators of heavy metals due to their filamentous morphology and high content of cell wall material [19]. The hyphal wall of fungi has been found to be the primary site of metal ion accumulation owing to the presence of the acetamide group of chitin, amino and phosphate groups in nucleic acids, amino, amido, sulfhydryl and carboxy groups in proteins, and hydroxyls in polysaccharides [20].

Influence of physico-chemical parameters on biosorption

Physico-chemical parameters affect the removal of thorium from aqueous solution, [14,21]. The fungal biomass



cell walls are formed of chitin, chitosan, glucan, lipid, phospholipids, which contain amino groups, carboxyl groups, phosphates, lipids, melanin, sulphates and hydroxides [22]. These functional groups are considered the sites sorption of the metals [23].

Study of the effect of physico-chemical parameters on the biosorption efficiency of thorium by the biosorbent reveals that biosorption percentage of Th(IV) ions follows a decreasing trend as metal ion concentration increased (Figure 3a). The biosorption of Th(IV) ions decreased from 80.48% to 70.45%

with an increase of Th(IV) ions concentration from 150 mg/L to 350 mg/L, however the biosorption efficiency by free biomass of *E. nigrum* reached to 54%. The maximum biosorption of Th(IV) ions was obtained at concentration of 150 mg/L. This behavior can be due to that, at higher concentrations, the increase of Th(IV) ions concentration probably lead to the increase in the number of ions competing for the available binding sites in the biomass resulted in the limitation of vacant binding sites, which saturate beyond certain concentrations, thus decreasing Th(IV) ions biosorption [24].

Furthermore, thorium biosorption percentage was strongly dependent on pH (Figure 3b). The capacity of functional groups found on the biomass cell wall surface depends on the pH of the solution [25,26]. Maximum removal of thorium was 83.54% as observed at pH 6.8 as shown in Figure 3b. At lower pH value, the cell wall of *E. nigrum* becomes positively charged and it is responsible for reduction in biosorption capacity. In contrary, at higher pH, i.e. more than pH 6.8,

the cell wall surface becomes more negatively charged and therefore the biosorption of Th(IV) onto *E. nigrum* is more due to attraction between the biomass and the positively charged metal ion. Under alkaline conditions, deprotonation of the heavy metals binding sites and precipitation of Th(IV) ions as insoluble hydroxides or hydrated oxides is believed to be responsible for the decrease in Th(IV) ions biosorption by *E. nigrum*.

Effect of amount of biosorbent on the biosorption of Th(IV) ions by *E. nigrum* is presented in Figure 3c. With an increase in biosorbent amount from 50 to 600 mg, the percentage of Th(IV) ions biosorption increased from 51.34% to 84.66%. Biosorption percentage of Th(IV) ions reached equilibrium at 500 mg biosorbent concentration. Increase in biosorption of Th(IV) ions due to the increase in biosorbent amount could be attributing to increase in surface area that leads to increase in the binding sites and combination of metal ion with the corresponding sites [27].

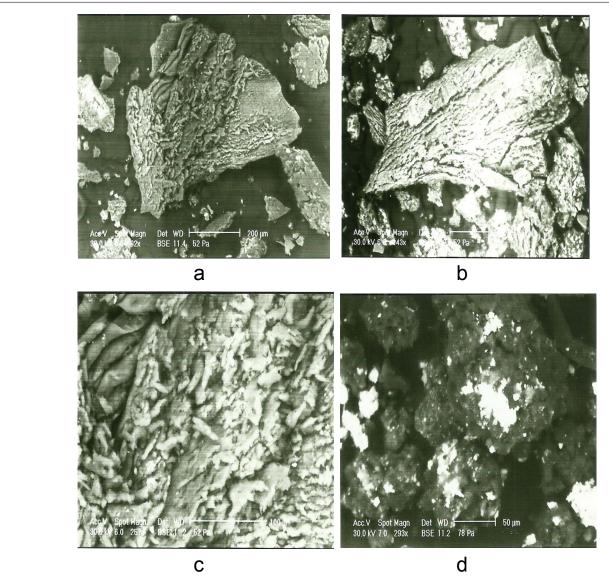
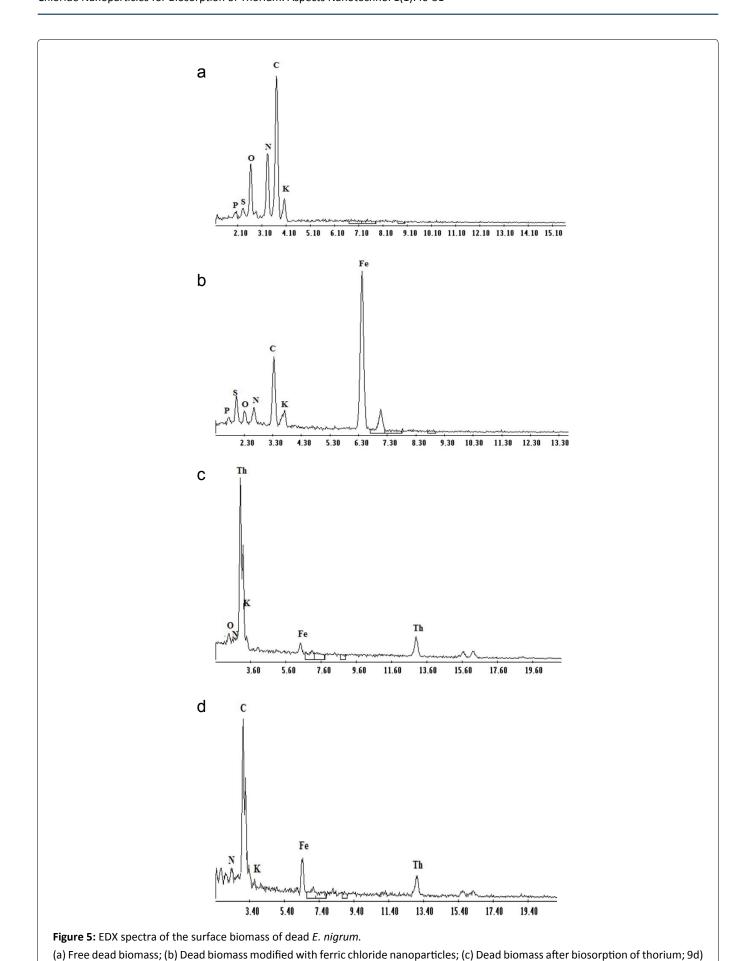


Figure 4: SEM micrograph of the surface biomass of dead *E. nigrum*.

- (a) Free dead biomass; (b) Dead biomass modified with ferric chloride nanoparticles; (c) Dead biomass after biosorption of thorium;
- (d) Dead biomass after elution by 0.1 N HCl.



Hussien et al. Aspects Nanotechnol 2018, 1(1):40-51

Dead biomass after elution by 0.1 N HCl.

Figure 3d presents Th(IV) ions biosorption, at different contact time, by E. nigrum. It was observed that there are two stages for the biosorption of Th(IV) ions; an initial rapid increase in biosorption percentage during the first 60 minutes, then followed by a plateau at 85.99% for Th(IV). This indicates that, as the contact time increased, more and more functional groups participated in biosorption of metal ions until it reaches equilibrium. Then when active sites on the biosorbents were filled, rate of biosorption became gradually constant and reached a plateau. In [14,16] reported similar trend of observation whereby the biosorption of nickel by Hypocrea lixii. The influence of temperature on thorium uptake is shown in Figure 3e. Maximum removal about 86.99% of Th (IV) occurred at 30 °C. Similar observation obtained in [28] as they reported that the increase in biosorption percentages with temperature may be attributed to either an increase in the number of active surface area sites available for biosorption or the decrease in the thickness of the boundary layer surrounding the biosorbent with temperature, so that the mass transfer resistance of adsorbate in the boundary layer decreases. The highest Th(IV) removal about 87.99% was observed at an agitation speed of 150 rpm as observed in Figure 3f. At high agitation speeds, the suspension is no longer homogenous, a fact impairing metal removal and biosorption process [29].

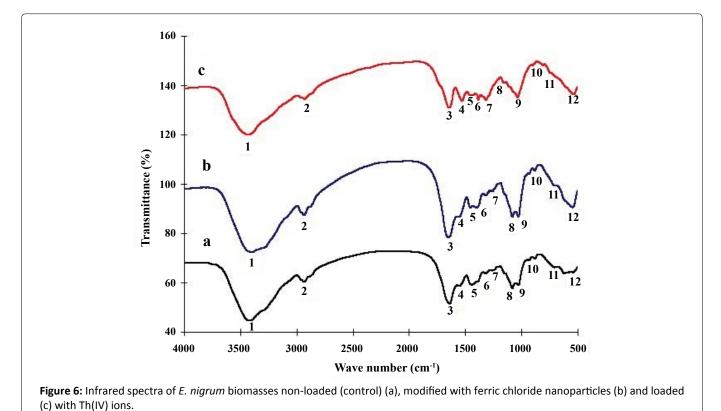
Characterization of pretreated *E. nigrum* biomass after biosorption of thorium

SEM observations: Thorium metals absorbed by the pretreated dead fungal biomass were characterized by morphological and structural analysis. Figure 4a, Figure 4b, and Figure 4c show SEM photomicrographs of free dead

biomass *E. nigrum* modified by ferric chloride nanoparticles before and after biosorption of Th(IV) ions by dead fungal biomass. Modification of the fungal surface area occurred due to formation of smooth characteristic surface as a glue stick coating the surface of the fungal biomass and leads to increasing in irregularity after binding to Th(IV) ions (i.e., Figure 4c). This result may be due to the biosorption affinity between the biomass and the element [30]. Figure 4d shows SEM photomicrograph of dead biomass of *E. nigrum* after elution by 0.1 N HCl. It was observed that the surface becomes clearer and a lot of biosorbed Th(IV) ions were eluted.

Figure 5a, Figure 5b, and Figure 5c illustrates The EDX spectra in the region of mycelium before and after thorium exposure. Figure 5c recorded in the examined region of mycelia indicates existence of signals of Th(IV) ions. However, signals of C, N and O were also observed, indicating the possible presence of proteins as a capping material. This protein matrix maybe responsible for the stabilization (capping material) of Th(IV) metals in the form of a film on the biomass surface. The obtained results agreed with that obtained in [16]. Also, EDX spectra in Figure 5d reveals that, 0.1 N HCl eluted Th(IV) ions with high efficiency reached about 80.87%. Table 1 indicates the Elemental analysis of the dead biomass of *E. nigrum* before and after thorium exposure.

IR spectral analysis: FTIR spectral analysis is important to identify some characteristic functional moieties which are responsible of adsorbing metal ions. Figure 6a, Figure 6b, and Figure 6c shows typical FTIR spectra for normal, treated *E. nigrum* biomasses non-loaded (control), pretreated with ferric chloride nanoparticles and loaded with Th(IV) ions respectively. FTIR spectra of the biosorbents are recorded in the range of 500-4000 cm⁻¹ to ascertain the information of



stretching and bending vibrations of the functional groups that are involved in the biosorption of metal ions.

According to FT-IR spectra, it is suggested that there is a physical and chemical attraction between Th(IV) ions and the binding sites on the *E. nigrum* biomass. This interaction

between Th(IV) ions and the sites on *E. nigrum* biomass modified the values of transmittance versus the wave number. Functional groups on *E. nigrum* have specific value of absorbance at different wave number (summarized in Table 1). Bonds at different wave numbers (3422, 2933, 1639,

Table 1: Elemental analysis of the dead biomass of *E. nigrum*.

	Dead biomass of <i>E. nigrum</i>	After modification with ferric chloride nanoparticles	After biosorption of thorium	After elution by 0.1 N HCl	
Р	1.85	7.44	UDL	UDL	
S	2.70	2.39	UDL	UDL	
0	14.33	3.44	UDL	UDL	
N	21.23	2.31	2.09	2.37	
С	55.88	14.87	1.13	80.87	
K	4.01	2.39	2.02	1.13	
Fe	UDL	67.16	4.33	12.39	
Th	UDL	UDL	90.43	3.25	
Wt%	100	99.97	100	100.1	

UDL: Under detection limit.

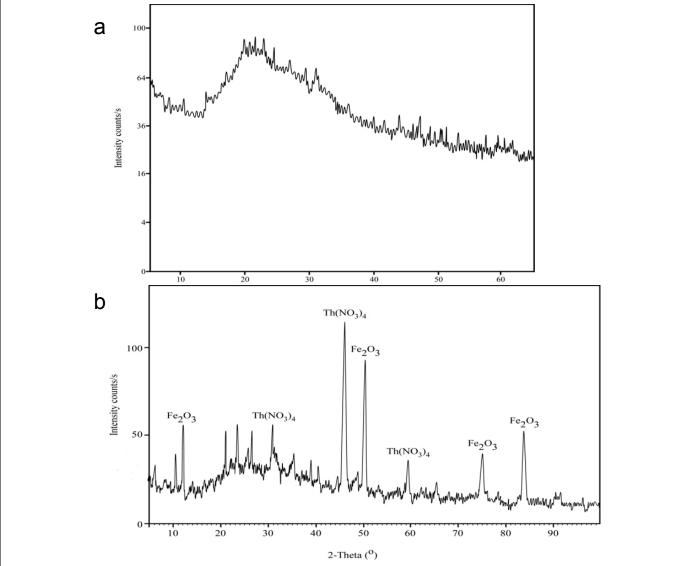


Figure 7: X-ray diffraction spectrum of (a) Control, modified biomass with ferric oxide nanoparticles; (b) Biomass after biosorption of thorium nitrate.

1549, 1439, 1319, 1260, 1081, 930, 880,707 and 540 cm⁻¹) on *E. nigrum* before biosorption were detected as O-H, C-H, C=C, N-O, O-H, C-N, C-O, C-O, =C-H, C-H and C-X, respectively.

Values of transmittance and wave number of bonds on the *E. nigrum* after modified with ferric chloride nanoparticle has been changed due to forming new bonds. The new bonds formed at different wave number of 1650, 1554 and 1452 cm⁻¹ correspond to the bending vibrations of the C-N of amide I, II and C=C of aromatic group, respectively (Figure 6 and Table 2). The amide linkages between amino acid and residues of proteins in cell wall of fungal stain give rise to well-known signatures in the infrared region of the electro-magnetic spectrum. Such mode arises from peptides/proteins bound to ferric nanoparticles which suggest the possibility of this agent acting as a carrier and capping agents [31].

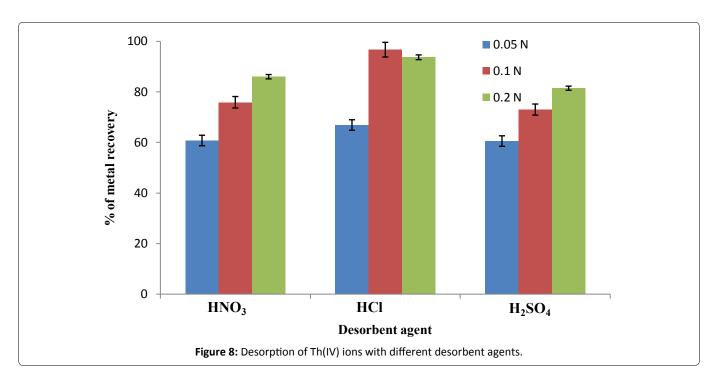
In this study, after saturating the biomass samples with Th(IV) ions, several bands shifts were observed in the FT-IR

spectra in relation to pretreated biomass, especially those assigned to amide groups. The bands at 3393 and 2933 cm⁻¹ were shifted to 3427 and 2926 cm⁻¹, respectively.

The bands at 1650 and 1554 cm⁻¹ were shifted to 1644 and 1528 cm⁻¹, respectively (Figure 6). It is suggested that biosorption process is due to the interaction between thorium ions and amide groups within the available binding sites on biomass. The bands observed at 1318 cm⁻¹ can be assigned to the C-N stretching vibrations of the aliphatic amines was shifted to 1310 cm⁻¹ (Figure 6) [32]. Such observations indicate the presence and binding of proteins to Th(IV) ions which can lead to their possible stabilization. In dead biomass probably the protein from the cell is liberated during the homogeneity process and bound on the surface cell. It was observed also appearance of band at 1385 cm⁻¹, and this can be assigned to the N-O stretching vibrations of nitro group and disappearance of band at 931 cm⁻¹ can be attributed to

Table 2: Functional groups of *E. nigrum* biomasses and the corresponding infrared absorption wave lengths.

Type of Biomass	Functional group	Type of compound and vibration	Frequency (cm ⁻¹)	Reference frequency range (cm ⁻¹)	Intensity
Before Biosorption	O-H	Hydrogen bonded alcohols & stretching	(1) 3422	3200-3600	Strong and broad
	C-H	Alkanes & stretching	(2) 2933	2850-3000	Strong
	C=C	Alkenes & stretching	(3) 1639	1648-1638	Variable
	N-O	Nitro & stretching	(4) 1549	1550-1500	Strong
	-COO-	Carboxylic acid & bending	(5) 1439	1449-1395	Medium
	C-N	Amine & stretching	(6) 1319	1080-1360	Medium
	C-O	Alkyl aryl ether & stretching	(7) 1260	1000-1300	Strong
	C-O	Primary alcohol & stretching	(8) 1081	1085-1050	Strong
	=C-H	Alkene & bending	(9) 931	675-1000	Variable
	C-H	Aromatic (meta-substituted benzene) & bending	(10) 880	860-900	Strong
	C-H	Aromatic (monosubstituted benzene) & bending		700-750	Strong
	C-X	Chloroalkanes & bending	(12) 540	540-760	Strong
	О-Н	Hydrogen bonded alcohols & stretching	(1) 3393	3200-3600	Strong and broad
	C-H	Alkynes	(2) 2933	2850-3000	Strong
	N-H	Amine-amide I & bending	(3) 1650	1650-1580	Medium
	N-H	Amine-amide II & bending	(4) 1554	1540-1640	Medium
	C=C	Aromatic & stretching	(5) 1452	1400-1600	Weak
After modified	C-N	Amine & stretching	(6) 1319	1080-1360	Medium
with FeCl ₃ nanoparticles	C-O	Alkyl aryl ether & stretching	(7) 1261	1000-1300	Strong
nanoparticles	C-O	Primary alcohol & stretching	(8) 1026	1085-1050	Strong
	=C-H	Alkene & bending	(9) 930	675-1000	Variable
	C-H	Aromatic (meta-substituted benzene) & bending	(10) 880	860-900	Strong
	C-H	Aromatic (monosubstituted benzene) & bending	(11) 707	700-750	Strong
	C-X	Chloroalkanes & bending	(13) 540	540-760	Strong
	О-Н	Hydrogen bonded alcohols & stretching	(1) 3427	3200-3600	Strong and broad
	C-H	Alkynes	(2) 2926	2850-3000	Strong
	N-H	Amine- amide I & bending	(3) 1644	1650-1580	Medium
	N-H	Amine- amide II & bending	(4) 1542	1540-1640	Medium
	C=C	Aromatic & stretching	(5) 1453	1400-1600	Weak
	N-O	Nitro & stretching	(6) 1381	1345-1385	Strong
Pretreated biomass	C-N	Amine & stretching	(7) 1310	1080-1360	Medium
	C-O	Alkyl aryl ether & stretching	(8) 1154	1000-1300	Strong
after Th(IV)	C-O	Primary alcohol & stretching	(9) 1032	1085-1050	Strong
biosorption	C-H	Aromatic (meta-substituted benzene) & bending	(10) 860	860-900	Strong
	C-H	Aromatic (monosubstituted benzene) & bending	(11) 740	700-750	Strong
	C-X	Chloroalkanes & bending	(12) 540	540-760	Strong



the =C-H bending vibration of alkenes.

The obvious explanation of a shift in the spectra, reduction and disappearance of the peaks are that there is an effect of chemical pretreatment and metal biosorption on the functional groups. These obtained results are in agreement with that obtained in [33].

X-ray diffraction (XRD): XRD analysis before and after Th(IV) biosorption are shown in Figure 7. In Figure 7a, there is a wide diffraction peak at the incident-beam 2θ of 14°-30°, which is formed by polysaccharide and other organic ingredients. The shape of this peak indicated that the original absorbents had an amorphous structure. In Figure 7b, obtained X-ray diffraction spectrum shows distinct peaks indicating the deposition of crystallized hematite and thorium ions. Diffractogram of the hematite displayed maximum number of peaks to iron between 2θ of 50.36° , 2θ of 77.35° and 20 of 84.49° as illustrated in Figure 7b. Significantly D values of most the peaks corresponding to iron chloride. Moreover, XRD pattern of Th(IV) loaded biomass Figure 7b is distinct and complex with some peaks narrowing and two new peaks emerging at 2θ of 32.04° , 2θ of 47.30° and 2θ of 60.35°, indicating deposition of thorium nitrate in the form of crystallized thorium. A similar result was obtained in [34].

Desorption study

In this section, desorption study was made to verify the possibility of desorbing the immobilized metal ions from the biomass. The percentage of desorbed Th(IV) ions from the biomass using hydrochloric, nitric and sulfuric acids at different concentrations for three washing cycles were shown in Figure 8. These desorbent agents were chosen because mineral acids appear to be the most feasibly applicable eluting agents. It was found that 0.1 N HCl desorbent agent eluted about 94.89% of the biosorbed Th(IV) ions by *E. nigrum* biomass after three cycles. Higher acid concentrations could be resulted in higher concentration of protons that swept the

metal ions away from the biosorbent. As expected, a slightly lower in weight loss could be achieved by 0.1 N HCl compared to other acid concentrations. so, it safe to be used for desorption purpose without increasing the biomass weight loss [10]. In [35] it is mentioned that the most common acid concentration of 0.1 M HCl desorbed about 96% of Th(IV) ions from loaded biomass, deoiled karanja seed cake (DKSC) an agro-industrial waste resulted after expelling oil from the seeds.

Thorium biosorption from (Th-U) concentrate solution

Chemical composition of (Th-U) concentrate solution obtained from Egyptian monazite was presented as it consists of 20.07% RE $_2$ O $_3$, 4.45% Fe $_2$ O $_3$, 19.01% ThO $_2$ and 2.44% UO $_2$. In addition to, CaO, Na $_2$ O, P $_2$ O $_5$ and loss of ignition (LOI) which presented in percentages 0.14%, 0.50%, 0.96% and 18.11%, respectively [36]. Under the previously obtained optimum conditions for biosorption of Th(IV) ions, it was found that the percentage of removal Th(IV) ions from (Th-U) concentrate solution was to be 78%. Furthermore, 87% of biosorbed Th(IV) ions were desorbed by 0.1 N HCl. Less removal of Th(IV) ions occurred in REEs solution as compared to synthetic Th(IV) ions solution might be attributed to the presence of interference of other rare earths in the solution that compete with Th(IV) ions for binding sites on the biosorbent. Similar observation was also reported in [37].

Conclusion

Dead biomass of the fungus *E. nigrum* was found to be an efficient tool for the uptake of thorium metal ions from aqueous solution. The dead biomass also acts as a stabilizer for the uptake of thorium ions during bioremediation processes. Th(IV) ions removal about 87.99% was observed after studying the optimum conditions. Results indicate that the presented novel approach for the bioremediation of

thorium ions is fast, low cost, environment friendly and easily scalable, using as reducing agent the fungus *E. nigrum*. SEM, XRD and IR spectra for *E. nigrum* biomass before and after adsorption of Th(IV) ions showed that different functional groups were involved in the biosorption of the metal ions. Furthermore, it was found that *E. nigrum* biosorbent uptake about 78% of Th(IV) ions from (Th-U) concentrate solution under the obtained optimum conditions. Removal of Th(IV) ions by *E. nigrum* biosorbent is in-line with alternative sustainable technology development for removal of Th(IV) ions from (Th-U) concentrate solution.

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